THE RELATIONSHIP OF FIELD BURN SEVERITY MEASURES TO SATELLITE-DERIVED BURNED AREA REFLECTANCE CLASSIFICATION (BARC) MAPS

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ABSTRACT

Preliminary results are presented from ongoing research on spatial variability of fire effects on soils and vegetation for from the Black Mountain Two and Cooney Ridge wildfires. Extensive field fractional cover data were sampled to assess the efficacy of quantitative satellite image-derived indicators of burn severity. The objective of this study was to compare the field burn severity measures to the digital numbers used to produce Burned Area Reflectance Classification (BARC) maps. Canopy density was the most highly correlated field variable to BARC data derived from either SPOT multispectral (XS) or Landsat Thematic Mapper (TM) imagery. Among ground cover variables, the depth of new litter, old litter and duff correlated better with the satellite data than did the fractional cover data (new and old litter, ash, soil and rock). Most field variables, with the notable exception of ash, tended to vary more at low and moderate severity sites than high severity sites. Semivariograms of the field variables revealed spatial autocorrelation across the spatial scales sampled (2 – 130 m), which the 20 m or 30 m resolution satellite imagery only weakly detected. In future analyses, we will quantify burn severity characteristics in other forest types and broaden our assessment to consider erosion processes, such as soil water penetrability following fire.

INTRODUCTION

The USDA Forest Service (USFS) and other land management agencies employ remote sensing tools to efficiently and effectively manage fire-adapted ecosystems. The USFS Remote Sensing Applications Center (RSAC) provides satellite imagery analysis for managing and monitoring wildfires. Burned Area Reflectance Classification (BARC) maps produced by RSAC have proved very useful for Burned Area Emergency Rehabilitation (BAER) teams during recent fire seasons; collaboration between RSAC and BAER will expand further in the upcoming fire seasons. One widely used indicator of burn severity is the Normalized Burn Ratio (NBR), calculated as the difference between the short-wave infrared and near-infrared channels, divided by their sum. Also used is the delta NBR (dNBR), calculated by subtracting post-fire from pre-fire NBR values (Key and Benson 2001). In a field validation of BARC products, Bobbe et al. (2003) found the dNBR to be no more accurate than NBR (Bobbe et al. 2003). Landsat Enhanced Thematic Mapper Plus (ETM+), Landsat Thematic Mapper (TM), and SPOT Multispectral (XS) imagery are the
satellite sensors most commonly used for mapping wildfires. However, failure of the scan line corrector on Landsat 7 (http://landsat7.usgs.gov/updates.php) has drastically compromised the provision of ETM+ imagery and forced greater reliance on Landat 5 TM and SPOT 4 XS imagery to map wildfires.

The two wildfires chosen for this analysis, Black Mountain Two and Cooney Ridge, were located on either side of Missoula, Montana (Figure 1). The Black Mountain Two wildfire ignited 11 Aug 2003 and burned 2855 ha before containment on 2 Sep 2003. The Cooney Ridge wildfire ignited 8 Aug 2003 and burned 8589 ha before containment on 14 Sep 2003. Fire severities ranged from none (some areas were unburned), to low, moderate and high. The fires occurred in complex terrain in mid to high-elevation mixed conifer forests in western Montana, USA. Areas burned most severely are relatively uniform, while the heterogeneity of moderate and low severity burns is substantially higher. On the Cooney Ridge fire, severity was observed to be more uniformly higher on privately owned lands that recently had been heavily logged (Stone et al. 2004).

The objectives of this Joint Fire Science Program (JFSP) project are to compare alternative remote sensors and analysis approaches across a diversity of soils, vegetation, and fire conditions, and to explicitly link fire behavior, fuels and fire effects to quantitative indicators of burn severity, that can be assessed in the field, predicted from fire effects models and mapped remotely.

**Figure 1.** Location of the Black Mountain Two and Cooney Ridge wildfires in western Montana.

**METHODS**

**Satellite Image Processing**

RSAC uses the NBR or, if pre-fire imagery is available, the dNBR, to produce BARC products. In addition to the
BARC classification image. RSAC provides an unclassified image of the BARC, called the BARC-Adjustable (BARCA). The BARCA product is simply the continuous NBR values stretched across the full dynamic range (0-255). Jenks Natural Breaks logic is used to assign temporary break points to the continuous BARCA variable and produce the BARC classification. The BARCA product allows BAER teams or other users to assign their own break points based on their own ground observations.

After classifying the BARC image, RSAC overlays the National Land Cover Database (NLCD) vegetation layer. Burned grasslands, for example, are often misclassified as “high” despite the fact that grasslands rarely burn severely, and so cause little concern for BAER teams. RSAC uses the NLCD GIS overlay to catch any grassland areas classified as “high” during the NBR process and reclassifies them as “low.” For the Black Mountain 2 and Cooney Ridge wildfires, the NLCD vegetation overlay led to no modifications of the BARC product.

Pre- and post-fire Landsat 5 TM images (10 Jul 2002 and 31 Aug 2003, respectively) were used to produce BARC and BARCA products from dNBR values for the Cooney Ridge fire. A post-fire SPOT 4 XS image (1 Sep 2003) was used to produce BARC and BARCA products from NBR values for both the Cooney Ridge and Black Mountain 2 fires (Figure 2). The Black Mountain 2 and Cooney Ridge wildfires were advantageous for evaluating Burned Area Reflectance Classification (BARC) products because the post-fire Landsat and SPOT images used to produce them were acquired at very nearly the same time. SPOT 4 has the advantage of being pointable, enabling simultaneous acquisition of the Black Mountain 2 and Cooney Ridge fires. Unfortunately, Landsat 5 is not pointable, so the Landsat-derived BARC and BARCA products include only the Cooney Ridge fire, since the Black Mountain 2 fire was situated just outside the image extent.

Field Measurements and Analysis

Field data for this analysis were gathered between 11 Sep and 13 Oct, 2003. Four sites were located at Black Mountain 2 (1 High, 1 Moderate, 2 Low) while six sites were located at Cooney Ridge (1 High, 2 Moderate, 3 Low). Sites were randomly located 80-300 m from the access road, in a large patch of what was viewed as a broadly representative severity type. Burned sites with predominantly green crowns were classified as low, with predominantly brown crowns as medium, and with predominantly black crowns as high. More low severity sites were sampled than
moderate, and more moderate than high, because of prior evidence that spatial heterogeneity in burn severity characteristics increases as burn severity decreases (Turner et al. 1999).

At each site, nine plots were situated in a 120 m x 120 m cross pattern, with the two legs bisecting each other in the form of a giant ‘X’. One leg was oriented along the prevailing slope and the other leg was oriented perpendicular to it, or across slope. A plot was situated at the end of each leg, 60 m from the center plot, with another plot situated in between by 20, 30 or 40 m. Each plot consisted of 15 subplots situated in 3 rows with 5 subplots/row; rows were spaced 4 m apart while subplots in the same row were spaced 2 m apart. Distances between plots were measured using a laser rangefinder, while subplots were laid out using a cloth measuring tape and marking each subplot with reusable pin flags.

Ocular fractional cover estimates of rock, mineral soil, ash, litter (new and old) and any large organics were made at each 1 m² subplot, with the visual aid of a 1 m x 1 m quadrat constructed from pvc pipe. Percent char of each cover component was also recorded. Four canopy density estimates were made from the center of each plot, facing the four cardinal directions, using a convex spherical densiometer. Plot centers were geolocated by logging a minimum of 150 positions with a Trimble GeoExplorer and subsequently differentially correcting them. Subplot positions were calculated based on their known distance and bearing from plot center.

BARC data at the subplot positions were extracted from Arc/Info GRID layers using ERDAS Imagine software. The preliminary data analysis and graphing reported in this paper was performed in R with the exception of the semivariograms, which were produced with WinGsLib software. A normal score transformation, which produces a standard Gaussian cumulative distribution with mean equal to zero and variance equal to one (Deutsch and Journel 1998), was used to normalize the data before generating semivariograms, enabling their comparison on the same plot.

RESULTS

Ten sites with 9 plots/site and 15 subplots/plot amounted to 1350 samples for comparing field and SPOT-XS BARC data. Because the Landsat-5 image was limited to the Cooney Ridge fire, only the 6 Cooney Ridge sites could be compared to the Landsat-5 BARC data, or a total of 810 samples. Our canopy density measurements were the best field data correlate to the BARCA values, for either sensor (Figure 3). Depth measurements of new litter (deposited post-fire), old litter (pre-fire but not consumed) and were better correlates than the fractional cover estimates of new and old litter (Table 1). The opposite signs of the correlation coefficients is simply a consequence of the SPOT-XS and Landsat-5 pre-classification BARCA values being scaled in opposite directions with regard to burn severity.

Table 1. Pearson Correlation Coefficients Between Selected Field Variables and Pre-Classification BARC Digital Numbers (0-255) Derived from SPOT-XS and Landsat-5 Imagery.

<table>
<thead>
<tr>
<th>Field Variable</th>
<th>SPOT-XS BARC Correlation</th>
<th>Landsat-5 BARC Correlation</th>
</tr>
</thead>
<tbody>
<tr>
<td>Canopy Density</td>
<td>0.8389</td>
<td>-0.8779</td>
</tr>
<tr>
<td>New litter depth</td>
<td>0.2446</td>
<td>-0.6260</td>
</tr>
<tr>
<td>Old litter depth</td>
<td>0.6282</td>
<td>-0.7974</td>
</tr>
<tr>
<td>Duff depth</td>
<td>0.5946</td>
<td>-0.7772</td>
</tr>
<tr>
<td>New litter</td>
<td>0.4539</td>
<td>-0.6090</td>
</tr>
<tr>
<td>Old litter</td>
<td>0.6282</td>
<td>-0.6474</td>
</tr>
<tr>
<td>Ash</td>
<td>0.0021</td>
<td>0.3444</td>
</tr>
<tr>
<td>Mineral soil</td>
<td>-0.3325</td>
<td>0.3995</td>
</tr>
<tr>
<td>Rock</td>
<td>-0.5185</td>
<td>0.3849</td>
</tr>
<tr>
<td>Inorganic (mineral soil + rock)</td>
<td>-0.5898</td>
<td>0.6091</td>
</tr>
</tbody>
</table>
The range of variability in new litter, old litter and duff depth measured at low and moderate severity sites was considerably broader than at high severity sites (Figure 4). A similar pattern was observed for the old litter cover fraction, while ash cover varied much more on high severity sites, and exposed soil and rock cover varied fairly consistently across all severities (Figure 5).
Figure 4. Boxplots of new litter, old litter and duff depths, by severity class (H=High, L=Low, M=Moderate).
At each site, the smallest lag distance separating subplots was 2 m while the largest was approximately 130 m, which defined the lower and upper distance thresholds of the sample semivariograms (Figure 6). The fairly high nugget semivariances of the field sample semivariograms was evidence of some finer scale autocorrelation (<2 m) that was uncaptured at the 2 m sampling frequency. The variable shape of the field variable semivariograms was evidence of considerable spatial variation at both the subplot (2-12 m) and plot (12-130 m) scales. That none of these semivariograms (with the possible exception of soil) had reached their sill was evidence of continuing spatial autocorrelation at longer lags. The reduced semivariance and smoother shape of the image-derived semivariograms was due to their much coarser sample resolution (20 – 30 m), yet they captured some spatial autocorrelation at the intermediate lags, as evidenced by their rounded shape.
Figure 6. Semivariograms of ash, rock, soil and organic cover fractions, and pre-classification BARC digital numbers derived from SPOT-XS and Landsat-5 imagery.
DISCUSSION

Most any type of remotely sensed imagery is more likely to capture fire effects on tree crowns than on the ground, because the canopy occludes the ground. Indeed, we found canopy density to be a better correlate to BARCA image values than any of the ground variables (Figure 3, Table 1). The better correlations between the satellite data and the new litter, old litter and duff depth data, in comparison to the new and old litter cover fractions, was surprising because we expected the cover data to have more influence on the satellite signal. Even so, the stronger relationship to the depth data is not evidence of causation (Figure 4). Mineral soil and rock behaved similarly in relation to the satellite data, which is why they were combined into a single inorganic cover fraction to present these results (Figure 5). Ash cover was perhaps the most unpredictable variable because it is redistributed by wind and water very quickly following fire. In future analyses, fire progression data will be extracted to measure time elapsed from burning until the data were acquired by the remote sensor or collected in the field. Other variables also come into play with time since burning, namely needlecast (new litter) and green vegetation regrowth. Bear grass and other vegetation had not yet resprouted at the time these satellite images were acquired and was a very minor factor when these sites were sampled in the field.

Higher semivariance in the field variables was expected due to their higher sampling frequency in comparison to 20 m or 30 m SPOT or Landsat pixels, respectively (Figure 6). Thus the satellite data was only suited for capturing spatial autocorrelation at the plot scale. We recently received hyperspectral imagery of finer spatial resolution (4 m) to assess burn severity patterns at the subplot scale. Spectral mixture analysis will be applied to the hyperspectral data to remotely estimate and map cover fractions of vegetation, litter, ash, mineral soil and rock, and char (Roberts et al. 1993). Semivariograms similar to those presented here will be generated from the discontinuous cover fractions estimated in the field and the continuous cover fractions estimated by spectral mixture analysis. These semivariograms will then be fit with appropriate mathematical models to enable geostatistical simulation of burn severity patterns (Dungan 1999). Better knowledge of burn severity patterns in relation to topographic variables should improve validation of fire behavior and spread models.

The few diagnostics presented here will be expanded to include other field variables, such as crown characteristics and soil water penetrability, and other remote sensing variables, such as fractional cover estimates to be derived from the hyperspectral imagery. We will expand our study beyond the two wildfires reported here to four other wildfires sampled in 2003, as well as 2004 wildfires. As a result of our research, we will identify burn severity indicators that are readily mapable and scalable – that is, measurable remotely as well as on the ground. Our project will improve the assessment of the severity of post-fire effects, including the potential for erosion and sedimentation, and thus the strategic effectiveness of post-fire rehabilitation.

REFERENCES


