

1 Abella et al.: Native species interactions with red brome

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3 **Relationships of Native Desert Plants with Red Brome (*Bromus rubens*): Towards**
4 **Identifying Invasion-Reducing Species**

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13 The interactions between native and exotic species occur on a continuum from facilitative to
14 competitive. A growing thrust in invasive species science is differentiating where particular
15 native species occur along this continuum, with practical implications for identifying species that
16 might reduce the invasibility of ecosystems. We used a greenhouse experiment to develop a
17 competitive hierarchy of 27 native species with red brome, an invasive annual grass in
18 southwestern USA arid lands, and a field study to assess *in situ* responses of brome to native
19 perennial species in the Mojave Desert. Native species most competitive with brome in the
20 competition experiment included the annuals pincushion and bristly fiddleneck and the
21 perennials eastern Mojave buckwheat, sweetbush, and brittlebush that reduced brome biomass to
22 49 to 70% of its grown-alone amount. There was no clear difference in competitive abilities
23 with brome between annual and perennial natives, and competitiveness was not strongly

24 correlated ($r = 0.15$) with the biomass of native species. In the field, sweetbush and brittlebush
25 supported among the least cover of brome, suggesting congruence of the strong early competitive
26 abilities of these species with *in situ* patterns of brome distribution. At the other extreme, brome
27 attained its highest average cover (19%) below littleleaf ratany, significantly greater than all but
28 three of the 16 species evaluated. Cover of brome was only weakly related ($r = 0.19$) to area of
29 the perennial canopy, suggesting that factors other than the sizes of perennial plants were linked
30 to differences in brome cover among species. Results suggest that (i) interactions with brome
31 differ substantially among native species, (ii) these interactions are not as closely linked to
32 biomass production as in more temperate regions, and (iii) there is potential for identifying
33 native species that can reduce invasion of desert ecosystems.

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35 **Nomenclature:** red brome, *Bromus rubens* L.; pincushion, *Chaenactis stevioides* Hook. & Arn.;
36 bristly fiddleneck, *Amsinckia tessellata* A. Gray; eastern Mojave buckwheat, *Eriogonum*
37 *fasciculatum* Benth.; sweetbush, *Bebbia juncea* (Benth.) Greene; brittlebush, *Encelia farinosa* A.
38 Gray ex Torr.; littleleaf ratany, *Krameria erecta* Willd. ex Schult.

39 **Key words:** Competition, facilitation, plant-plant interactions, invasion-reducing communities,
40 invasibility, revegetation.

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47 Interactions between plants occur on a continuum. This continuum ranges from
48 facilitation, where fitness of one or both species is enhanced, to competition, where the fitness of
49 one or both is reduced (Keddy 2001; Brooker et al. 2008). Recently, interest in these interactions
50 has grown in invasive species science for identifying native species and communities that are
51 most competitive with exotic species and capable of reducing invasion (Levine et al. 2004;
52 Thacker et al. 2009). If competition from native species is to effectively suppress exotic species,
53 the native species must usurp resources (e.g., soil water or nitrogen) required by an exotic,
54 occupy space (e.g., regeneration microsites) otherwise inhabitable by an exotic, exude
55 allelopathic chemicals or negatively modify sites, interfere with reproduction or emergence from
56 the soil seed bank, or reduce fitness through other means (Seabloom et al. 2003). For these
57 purposes, competition is broadly defined as any reduction in the fitness of exotic species due to
58 the presence of native species (Fowler 1986).

59 Several experiments have demonstrated that species composition of resident communities
60 is a key factor affecting the competitive (or facilitative) environment to which a potential invader
61 is exposed (Brown et al. 2008). Species differ in traits such as height, biomass, growth rate, and
62 root:shoot ratio that are related to a species' ability to capture resources and are linked to
63 competitive ability (Booth et al. 2003). The particular traits that confer competitive or
64 facilitative advantages across different communities (e.g., arid versus temperate ecosystems) are
65 generally not well articulated, making it difficult to accurately predict which native species might
66 be the most competitive with exotic species (Barney et al. 2005). Assessing species interactions
67 in trait- and species-rich communities also requires many pair-wise comparisons, making it
68 difficult to test more than a few species. One approach to screening the competitive abilities of
69 many species is to grow all species of interest with the same species, termed a "phytometer"

70 (Keddy 2001). Keddy et al. (2002) used this approach to compare the competitive abilities of 63
71 herbaceous species from temperate eastern Ontario. Competitive abilities were hierarchical, with
72 native species reducing phytometer biomass by 45 to 99%. A potential disadvantage of using a
73 phytometer to screen competitive ability is that competition is only determined against the
74 phytometer species. However, this is a strength precisely when competitive ability against a
75 particular species is of interest, such as native species versus a target exotic species.

76 Exotic grasses such as red brome (*Bromus rubens* L., hereafter *Bromus*) infest extensive
77 areas of southwestern arid lands including the Mojave Desert. These exotic species have
78 increased fuel loads, facilitating widespread fires in desert ecosystems not thought to have
79 burned extensively prior to the widespread establishment of these species by the mid-1950s (Salo
80 2005). For example, more than 2% of the entire area of the Mojave Desert burned in 2005 alone
81 following a moist year that stimulated vigorous annual plant growth (Brooks and Matchett 2006).
82 These fires kill late-successional perennial vegetation, often reducing perennial cover for ≥ 40
83 years and altering native species composition for indeterminate time periods (Abella 2010). The
84 extensiveness of invasion across millions of hectares in the arid West makes reducing dominance
85 of exotic grasses a challenging task confronting resource managers. A complicating factor is that
86 some native perennial species may actually facilitate the establishment and growth of exotic
87 plants by forming fertile islands high in soil nutrients and with ameliorated microclimates
88 (Brooks 2009; Craig et al. 2010).

89 It is not well known where different native annual and perennial species fall on the
90 continuum of facilitative to competitive with exotics such as *Bromus* in southwestern arid lands,
91 or if competition from native species is an effective management tool for reducing exotic plants
92 (Abella and Newton 2009). The objectives of this study were to: develop a *Bromus* competitive

93 hierarchy under controlled greenhouse conditions with a range of species native to the Mojave
94 Desert, and to assess *in situ* patterns of *Bromus* cover below native perennial species. We posed
95 the following questions: (1) Can native desert annual and perennial species be arranged in a
96 competitive hierarchy with *Bromus*? (2) Are interactions of native species with *Bromus*
97 correlated with species traits such as root mass? (3) Does *Bromus* cover vary in the field among
98 native perennial species, and do these results support those of the competition experiment? We
99 hypothesized that all native species would reduce the growth of *Bromus* in the competition
100 experiment, but the relative reduction would vary among species. Based on studies such as
101 Keddy et al. (2002), we further hypothesized that the strongest competitors with *Bromus* are
102 species that grow the most rapidly and produce the greatest biomass. We suspected that, owing
103 to rapid biomass production, annuals as a group would be more competitive than perennials. We
104 also anticipated that the poorest perennial competitors at the seedling stage in the competition
105 experiment would be the most heavily infested by *Bromus* in the field.

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Materials and Methods

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Autecology of the Study Species. *Bromus* is native to southern Europe, northern Africa, and southwestern Asia, occurring from sea level to 1,300 m in arid to mesic shrubland and steppe habitats (Brooks 2000). *Bromus* readily invades disturbed habitats in arid southwestern North America, but it also infests relatively undisturbed late-successional communities such as those containing creosote bush (*Larrea tridentata* (DC.) Coville) and blackbrush (*Coleogyne ramosissima* Torr.; Salo 2005). In southwestern arid lands, *Bromus* forms prodigious soil seed banks that can exceed 3,000 seeds/m² (0 to 5 cm soil depth) and are readily germinable given

116 moisture and warm spring temperatures (Abella et al. 2009a). *Bromus* often is most abundant
117 below the canopies of native perennial plants, although it can dominate interspaces during wet
118 years and on sites with favorable soils (Brooks 1999). In the Mojave Desert, for example,
119 Brooks (2003) reported that the aboveground biomass of *Bromus* in interspaces averaged $< 0.5 \text{ g}$
120 m^{-2} in 1996 (a dry year) and 5 g m^{-2} in 1997 (a moister year). In contrast, biomass below
121 perennials was 10 g m^{-2} in 1996 and 55 g m^{-2} in 1997, approximately 10- to 20-fold greater than
122 in interspaces. *Bromus* is considered one of the most abundant and widespread annual species
123 across the Mojave Desert (Beatley 1966; Hunter 1991; Brooks 2009).

124

125 **Study Area.** Our work focused on species inhabiting the Mojave Desert, a 124,000-km² hot
126 desert of southeastern California, southern Nevada, southwestern Utah, and western Arizona in
127 the southwestern USA (Rundel and Gibson 1996). A weather station (Las Vegas, NV, elevation
128 662 m) typifying the eastern Mojave Desert has reported averages of 11 cm yr^{-1} of precipitation,
129 July daily maximum temperature of 40°C , and January daily minimum temperature of 1°C (1937
130 to 2009 records; Western Regional Climate Center, Reno, NV). This desert receives $> 60\%$ of
131 its rainfall in winter, facilitating peak biomass of both native and exotic annual plants in late
132 spring (March through May) during moist years (Beatley 1974). Seedlings of many native
133 perennials also germinate during this period, with growth of the seedlings and mature plants
134 occurring primarily during the warm spring and summer months (Turner and Randall 1987).
135 The typical physiognomy of the desert is widely spaced native perennial plants, such as *Larrea*
136 *tridentata* and white bursage (*Ambrosia dumosa* (A. Gray) Payne), with interspaces between the
137 shrubs containing annual plants (in moist years) or sparsely vegetated soil.

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139 **Competitive Hierarchy Experiment.** Following Keddy et al. (2002) for developing competitive
140 hierarchies, we selected 35 species native to the Mojave Desert that ranged widely in growth
141 form (16 annual forbs, 3 perennial grasses, 7 perennial forb-subshrubs, and 9 perennial shrubs),
142 longevity (16 annuals and 19 perennials), and successional affiliation based on Abella (2010).
143 Nomenclature and growth form and longevity classification follow NRCS (2010). We obtained
144 seed^{1,2} of the native species from the plant nurseries of Lake Mead National Recreation and the
145 Las Vegas Springs Preserve, both within the Mojave Desert. We collected *Bromus* seed from
146 two populations within Lake Mead National Recreation Area. We conducted a pilot emergence
147 assay (one month duration, with 20 seeds sown in one 3.8-L pot per species with the same soil to
148 be used in the experiment) to determine seed emergence rates to estimate the number of seeds
149 needed to be sown to produce the desired density for the competition experiment.

150 Based on the emergence information, we sowed seeds of native species for the
151 competition experiment in separate pots to attain a density of 5 seedlings per pot (278 m⁻²). We
152 simultaneously sowed seeds of the *Bromus* phytometer in the same pots containing the native
153 species to also attain a density of 5 seedlings per pot for *Bromus*. Due to the large number of
154 species tested, we were only able to assess a single density. This density, however, is within the
155 range reported for Mojave Desert field sites in years that have precipitation patterns supporting
156 *Bromus* growth (Beatley 1966; Hunter 1991). To try and ensure close contact of natives with
157 *Bromus*, we used 3.8-L pots that were circular with a diameter of 15 cm. We filled pots with a
158 sterile soil mix consisting of 2:1 sand:potting soil³ to simulate the sandy loam textures within
159 ranges commonly characterizing Mojave Desert soils (Lato 2006). Our target was to have three
160 replicate pots for each native species, so we sowed seeds in four pots in case some pots lacked
161 sufficient seed germination to attain the target density. We also grew 5 seedlings per pot of

162 brome alone in 10 replicate pots. We randomly arranged pots on a bench in a greenhouse
163 maintained at 24°C without supplemental lighting. Pots were watered with a misting system in
164 operation for two minutes each day that delivered 1.5 cm of water per day.

165 We initiated the experiment in April 2009, when *Bromus* actively grows in the field
166 (Beatley 1966), and ran the experiment for eight weeks. Each week, we measured the heights of
167 native and *Bromus* plants and counted the number of plants per pot, thinning plants to the target
168 density of 5 plants per pot as needed. At the end of the experiment, which was the time of
169 senescence of *Bromus* and the native annuals, we harvested above- and below-ground biomass
170 by species within pots and determined oven-dry weight by heating at 70°C for 24 h. Since we
171 averaged data from all plants by species within pots, distinguishing root masses (which could
172 become intertwined by growing in close contact) of individual plants within species was not
173 necessary.

174 Despite the pilot study and oversowing many species, eight species failed to attain the
175 target density in at least three replicate pots and were not further included in the experiment.
176 While it is possible that manipulating seeds could have increased germination and emergence,
177 we chose to minimally manipulate seeds and avoid attempting species-specific treatments that
178 could have confounded among-species comparisons of competitive abilities. For example,
179 gibberellic acid may have enhanced germination of some species, but this treatment can affect
180 subsequent seedling growth (Baskin and Baskin 2001). A total of 27 species, 12 annuals and 15
181 perennials, had sufficient emergence for inclusion in the experiment.

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183 **Field Patterns of *Bromus* below Native Perennials.** We used seven sites, located in Lake Mead
184 National Recreation Area, Red Rock Canyon National Conservation Area, and surrounding

185 Bureau of Land Management land in southern Nevada in the eastern Mojave Desert, to assess the
186 cover of *Bromus* below native perennial species. We selected these sites because they contained
187 a variety of perennial species to permit within-site comparisons and harbored some of the species
188 included in the competition experiment. These sites also contained *Bromus* infestations, which
189 we identified based on a landscape-scale assessment of invasive species distributions in the study
190 area (Abella et al. 2009b). Elevations of the sites ranged from 710 to 1,367 m. The sites were in
191 the 421, 731, 732, and 750 mapping units of the Clark County soil survey and were primarily
192 classified as loamy-skeletal, carbonatic, mesic, shallow Calcic Petrocalcids and Typic and Lithic
193 Torriorthents (Lato 2006). These soils have gravelly loam and sandy loam textures (Lato 2006).
194 Sites had a geographic extent of 64 km, spanning the Universal Transverse Mercator coordinates
195 of 642,316 m E, 4,001,353 m N and 702,253 m E, 3,981,227 m N (zone 11, North American
196 Datum 1983).

197 At the center of each site within an area of 1 ha, we established four belt transects (one in
198 each of the cardinal directions) that were 5 m wide and up to 50 m long. Along each transect, we
199 measured the first two individuals of any perennial species intended to be included in the
200 competition experiment. We required that an individual have a developed canopy (which
201 excluded sampling seedlings of these species) that did not overlap with the canopy of any other
202 perennial species. We measured the perimeter of each individual canopy (at its greatest extent)
203 and calculated canopy area from perimeter using the formula for a circle to approximate the
204 shape of the canopies. In a vertical plane to the ground below canopies, we visually categorized
205 areal cover of *Bromus* below each perennial plant using the following cover classes: 0 = 0%
206 cover, 1 = > 0 < 1%, 2 = 1 to 3%, 3 = 3 to 6%, 4 = 6 to 12%, 5 = 12 to 25%, 6 = 25 to 50%, 7 =
207 50 to 75%, and 8 = 75 to 100%. Two interspaces (defined as an area ≥ 1 m from the nearest

208 perennial plant canopy) were also identified using random coordinates within each transect and
209 sampled in the same manner. The dimensions to be sampled within the interspaces were
210 determined based on the dimensions of two randomly selected perennials measured in transects.
211 *Bromus* was the predominant standing annual at all sites, although at one site small amounts of
212 the exotic annual grasses *Schismus* spp. and cheatgrass (*Bromus tectorum* L.) were intermingled
213 with the *Bromus* stalks. These species were included in the *Bromus* cover estimates, as the mat
214 of plant matter at this site was difficult to partition into the small amounts of the other species.
215 Because of *Bromus*'s contribution to hazardous fuels, standing dead biomass is as important as
216 green biomass. We made measurements from November 2009 to January 2010 when *Bromus*
217 was not actively growing and sampled standing dead biomass as a measure of recent
218 accumulated biomass. Dead *Bromus* plants can persist for more than two years (Beatley 1966).

219

220 **Data Analysis.** For the competition experiment, we calculated native species performance versus
221 *Bromus* as the percent of grown-alone biomass that *Bromus* attained when grown with a native
222 species (Equation 1).

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224 $(\text{Bromus biomass when grown with a native species} / \text{Bromus grown-alone biomass}) \times 100$ [1]

225

226 Because data did not meet assumptions of homogeneity of variance and normality, we compared
227 performance among species using a nonparametric Kruskal-Wallis test followed by Tukey's test
228 ($\alpha = 0.05$) for multiple comparisons in SAS software (PROC NPAR1WAY, followed by PROC
229 MIXED on the ranks; SAS Institute 1999). Since the original data were continuous, we report
230 means and standard errors of the means for interpretive purposes but present the nonparametric

231 statistical results. We used Pearson correlation coefficients, as Spearman rank coefficients were
232 similar, to examine relationships among *Bromus* and native species traits. Traits included root
233 and shoot length, and root, shoot, and total biomass for mono-specific and mixed-species pots.
234 All values in the experiment were averaged on a pot basis and expressed as averages per plant,
235 and were further averaged on a native species basis for analyses at the among-species level.

236 Despite selecting mixed-species sites for the field study, no site contained all of the
237 species that were included in the competition experiment and species composition of the sites
238 was not identical. Because of this, we viewed the sites and transects as frameworks in which to
239 sample within a defined area, and each perennial plant as a discrete individual unit for analysis.
240 In a non-parametric analysis to accommodate the categorical cover data, we compared the
241 median percent cover (based on midpoints of cover classes) among perennial species and
242 interspaces with a Kruskal-Wallis test followed by Tukey's test for multiple comparisons (SAS
243 Institute 1999). Confidence intervals at the 95% level were computed through the PROC
244 UNIVARIATE and CIQUANTDF option routine.

245

246 **Results**

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248 **Competition Experiment.** *Bromus* root, shoot, and total plant mass were strongly correlated,
249 with correlation coefficients (Pearson r) ranging from 0.84 to 0.97 when *Bromus* was grown with
250 native species. Root length and root mass ($r = 0.75$) and shoot length and shoot mass ($r = 0.85$)
251 also were strongly correlated. Based on these relationships, we assessed competitive effects of
252 natives on *Bromus* using *Bromus* total plant (root + shoot) biomass.

253 There was an overall effect of native species identity on *Bromus* biomass indicated by a
254 significant Kruskal-Wallis test (χ^2 statistic = 41.9; $P = 0.0336$), but P -values were > 0.06 in
255 pairwise multiple comparisons among species. We arranged species hierarchically based on
256 their mean effects on *Bromus* biomass. The strongest native annual competitors versus *Bromus*
257 included pincushion (*Chaenactis stevioides* Hook. & Arn.), bristly fiddleneck (*Amsinckia*
258 *tessellata* A. Gray), cleftleaf wildheliotrope (*Phacelia crenulata* Torr. ex S. Watson), and
259 desertbells (*Phacelia campanularia* A. Gray), which reduced *Bromus* biomass to 50 to 78% of its
260 grown-alone amount (Fig. 1). By reducing *Bromus* biomass to 49 to 70% of grown-alone
261 amount, the strongest native perennial competitors were eastern Mojave buckwheat (*Eriogonum*
262 *fasciculatum* Benth.), sweetbush (*Bebbia juncea* (Benth.) Greene), and brittlebush (*Encelia*
263 *farinosa* A. Gray ex Torr.). Several species, such as the annual California poppy (*Eschscholzia*
264 *californica* Cham.) and perennials desert marigold (*Baileya multiradiata* Harv. & A. Gray ex A.
265 Gray), big galleta (*Pleuraphis rigida* Thurb.), and brownplume wirelettuce (*Stephanomeria*
266 *pauciflora* (Torr.) A. Nelson), promoted greater *Bromus* biomass than when *Bromus* was grown
267 alone. The average competitive performance of native species was not strongly correlated ($r =$
268 0.15) with the mean total biomass per plant (root + shoot) they attained. The correlation also
269 was weak ($r = 0.24$) when considered on a pot basis rather than averaging across native species.

270

271 ***Bromus* Cover below Native Perennials.** A total of 56 interspaces and 319 individuals of 16
272 species were measured. The median cover of *Bromus* below perennial plants differed
273 significantly among species and interspaces (Kruskal-Wallis $\chi^2 = 144$; $P < 0.0001$). Interspaces
274 contained the lowest median *Bromus* cover, significantly lower at $P < 0.05$ than 10 of 16 (63%)
275 of the perennial species studied (Table 1). Species were arranged hierarchically with respect to

276 *Bromus* cover, with turpentinebroom (*Thamnosma montana* Torr. & Frém.), *Bebbia juncea*,
277 Virgin River brittlebush (*Encelia virginensis* A. Nelson), Mexican bladdersage (*Salazaria*
278 *mexicana* Torr.), and *Encelia farinosa* exhibiting lower tier medians. Littleleaf ratany (*Krameria*
279 *erecta* Willd. ex Schult.), Torrey's jointfir (*Ephedra torreyana* S. Watson), *Larrea tridentata*,
280 and cheesebush (*Hymenoclea salsola* Torr. & A. Gray) harbored among the greatest *Bromus*
281 cover among species. *Krameria erecta* had a median twice as high as the next highest species
282 and supported significantly greater *Bromus* cover than all but three of the other species.

283 *Bromus* cover was not strongly correlated with the area of a perennial plant canopy ($r =$
284 0.19) when all 319 individual plants were included. *Bromus* cover-perennial canopy area
285 correlations also were not strong within species. For example, the correlation was only 0.14 for
286 *Krameria erecta* and 0.16 for *Encelia farinosa*.

287

288 **Discussion**

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290 **Evidence for Hypotheses.** In contrast to our expectation that all native species would reduce
291 *Bromus* biomass by some extent in the competition experiment, *Bromus* actually exhibited
292 greater amounts of biomass (than when grown alone) when grown with species such as
293 *Stephanomeria pauciflora*, *Pleuraphis rigida*, and *Eschscholzia californica* (Fig. 1). These
294 results contrast with those of Keddy et al.'s (2002) competitive hierarchy experiment with
295 temperate eastern North American species, where all 63 species reduced biomass of the native
296 annual phytometer, fluxweed (*Trichostema brachiatum* L.). Since experimental designs were
297 similar between the studies, our contrasting findings could result from the exotic versus native
298 phytometer or reflect that desert species are not as strongly competitive as temperate species.

299 Also contrasting with our hypothesis, biomass production and plant size of native species
300 was not strongly related to competitive ability in the greenhouse or to the amount of *Bromus*
301 cover the native species supported in the field. Competitive ability in temperate regions
302 frequently is related to plant size (Keddy 2001). Keddy et al. (2002), for instance, found that
303 total and above- or below-ground biomass were most strongly correlated ($r = 0.53$ to 0.58) with
304 competitive performance. Plant size in our study also was not linked to facilitation of *Bromus* in
305 the field, as species such as *Krameria erecta* supporting the greatest *Bromus* cover were not
306 necessarily the largest. Similarly, competitive performances of annual and perennial species
307 overlapped (Fig. 1), incongruent with our supposition that annuals would outperform seedlings
308 of perennial plants. Other traits than plant size and annual/perennial status apparently are more
309 closely connected to outcomes of interactions with *Bromus*.

310 The results of the correlative field study of *Bromus* cover below perennial species
311 followed those of the greenhouse competition experiment more closely for some species than
312 others. *Eriogonum fasciculatum*, the top competitor with *Bromus* in the greenhouse, harbored
313 varying amounts of *Bromus* in the field to exhibit a middle ranking among species (Table 1).
314 Considering the next three top competitors in the greenhouse, *Bebbia juncea* and *Salazaria*
315 *mexicana* supported lower tier amounts of *Bromus* cover in the field, and *Encelia farinosa* low-
316 middle amounts. *Larrea tridentata*, *Ambrosia dumosa*, and *Hymenoclea salsola* ranked in the
317 middle in greenhouse competitive performance, and supported middle-high amounts of *Bromus*
318 in the field. *Pleuraphis rigida*, the least competitive species in the greenhouse that also was part
319 of the field study, ranked in the middle among species in the field. A longer term field
320 experiment tracking interactions with *Bromus* of planted perennials from the seedling through

321 adult stages could be useful for evaluating how competitive interactions might change through
322 species' life cycles.

323

324 **Comparing Species Performance.** Brooks (2009) is the only other study known to us that has
325 compared relationships in the native species versus *Bromus* direction (as opposed to *Bromus*
326 effects on natives) among an array of species in southwestern deserts. Brooks (2009) compared
327 *Bromus* biomass below 29 native perennial species in the southwestern Mojave Desert. While
328 not all species were the same between Brooks (2009) and our study, the large numbers of species
329 included in the studies enabled a relative ranking of species relationships with *Bromus*. We were
330 able to compare 10 species in Brooks (2009) that were included in our field study, six of which
331 also were included in our greenhouse experiment. Within the studies, we ranked these 10
332 species relative to all species included in a respective study (27 in our greenhouse experiment, 16
333 in our field study, and 29 in Brooks [2009]) into lower, middle, and upper thirds in their
334 performance versus *Bromus*.

335 Eight of the 10 species were within one performance category among studies, while two
336 species differed substantially (Table 2). For example, *Bebbia juncea* consistently ranked as
337 'good' for both reducing *Bromus* in the greenhouse and supporting low amounts of *Bromus*
338 cover in our and the Brooks (2009) field studies. *Eriogonum fasciculatum* ranked as 'good' in
339 the greenhouse and 'medium' in the two field studies, and *Ambrosia dumosa* consistently ranked
340 as 'medium'. Species that ranked 'poor' in our field study, such as *Krameria erecta*, ranked no
341 better in Brooks (2009) than 'medium'. *Salazaria mexicana* and *Thamnosma montana* were the
342 most disparate species, ranking 'good' in our study, but 'poor' in Brooks (2009).

343 Variability was high in performance among individuals within a species in both our
344 studies and Brooks (2009), where standard errors of the means were as large or nearly as large as
345 means for several species. Brooks (2009) highlighted two major spatial scales of variability in
346 native species relationships with *Bromus*: broad landscape scales such as elevation gradients,
347 where *Bromus* is typically most abundant at middle elevations, and within-site variation where
348 microsites such as interspaces and canopies of perennial plants are important. He further noted
349 that *Bromus*-native species relationships can shift with elevation. For example, *Bromus* may be
350 closely associated with *Larrea tridentata* at lower elevations, but as abundant or more abundant
351 in interspaces or below perennials other than *Larrea* at cooler and moister higher elevations.
352 Further evaluating the potential interactions of landscape gradients (e.g., elevation) with
353 perennial plants in affecting the distribution of *Bromus* could be useful future research.

354

355 **Species Traits and Performance.** While we did not find that the traits of biomass production in
356 the greenhouse experiment or plant size in the field study were correlated with native species
357 performance versus *Bromus*, other traits could be important. Some of the many traits of potential
358 importance for perennials could include: seasonality of growth, canopy morphology, life span,
359 degree and nature of fertile island formation, successional affiliation, litterfall and decomposition,
360 allelopathy, and rooting habit/mycorrhizal associations. For example, in a Mojave Desert field
361 experiment, James et al. (2006) reported that the seasonality of growth of the native shadscale
362 saltbush (*Atriplex confertifolia* (Torr. & Frém.) S. Watson) more closely corresponded with that
363 of the exotic annual Arabian schismus (*Schismus arabicus* Nees) and was a better competitor
364 with *Schismus* than the perennial Parry's saltbush (*Atriplex parryi* S. Watson). De Soyza et al.
365 (1997) found that differences in canopy morphology among *Larrea tridentata* individuals

366 influenced microclimates and associated native annual communities. This observation suggests
367 that differences in canopy morphology both within and among species could influence the
368 growing environment for *Bromus*. Based on observing distributions of native annuals in the
369 Mojave Desert, Muller (1953) suggested that perennials with shorter life spans harbored fewer
370 annuals. This supposition appears consistent with our field data for some species but not others.
371 For instance, *Bebbia juncea* supported little *Bromus* cover and was classified by Bowers et al.
372 (1997) as a short-lived perennial with an estimated maximum life span of 20 years. However,
373 *Eriogonum fasciculatum* and broom snakeweed (*Gutierrezia sarothrae* (Pursh) Britton & Rusby)
374 also are considered short lived with estimated life spans of 20 years, but these species harbored
375 intermediate amounts of *Bromus*. In addition to other traits, life span could influence the degree
376 and nature of fertile island formation (Holmgren et al. 1997). Thompson et al. (2005) showed
377 that soil properties such as nitrogen mineralization differed below canopies of three late-
378 successional shrub species in the northern Mojave Desert. It could be hypothesized that species
379 with shorter life spans, often with early successional affiliations (Bowers et al. 1997), do not
380 form well-developed fertile islands and hence would be less favorable for *Bromus* growth. The
381 five species that harbored the least amount of *Bromus* in our field study (Table 1) all are
382 considered as early successional based on Abella's (2010) successional classification of Mojave
383 Desert species. With the exception of *Hymenoclea salsola*, species that supported the greatest
384 cover of *Bromus*, such as *Krameria erecta* and *L. tridentata*, are classified as late successional.

385 Litterfall, allelopathy, and rooting habit/mycorrhizal associations could further influence
386 the growing environment for *Bromus* near perennial plants. Strojan et al. (1979) reported that
387 yearly litterfall amounts in grams per shrub varied by an order of magnitude among six shrub
388 species in the northern Mojave Desert. Depending on decomposition rates, this variation in

389 litterfall could result in different thicknesses of litter layers below shrubs, which could positively
390 or negatively or not influence *Bromus* establishment. Gray and Bonner (1948) isolated an
391 allelopathic growth inhibitor in the foliage of *Encelia farinosa*, which was observed to harbor
392 few annual plants. However, Muller (1953) subsequently found that an even stronger growth
393 inhibitor was present in *Ambrosia dumosa*, a species that harbored abundant annuals. The
394 significance of allelopathy as a factor influencing annuals below perennials remains unclear.
395 The possible role of rooting habit also is unclear. Wallace et al. (1980) reported that *Krameria*
396 *erecta* had more than twice the proportion of fine roots (of its total root system) in the upper 10
397 cm of soil among nine perennial species in the northern Mojave Desert. It could be hypothesized
398 that this shallow rooting intensifies competition against shallow-rooted annuals, yet *K. erecta*
399 harbored the greatest cover of *Bromus* in our field study (Table 1). Since *K. erecta* also can be a
400 root parasite on other perennial plants (Griffith 1991), which could influence nutrient contents in
401 soils below its canopy, these observations suggest that exploring the possible influence of rooting
402 ecology in *Bromus*-native species relationships may be a useful topic for future research.
403 Similarly, Titus et al. (2002) reported that *Baileya multiradiata* had the greatest concentration of
404 arbuscular mycorrhizae among 15 Mojave species. *Baileya multiradiata* was one of the species
405 that facilitated *Bromus* growth in the greenhouse experiment (Fig. 1), and it is unclear if
406 mycorrhizal associations influence interactions with *Bromus*.

407 While traits such as fertile island formation may be less important in native annual
408 species interactions with *Bromus*, other traits such as seasonality of growth, water requirements,
409 and resource use discussed above for perennials may also be important for annuals (e.g., DeFalco
410 et al. 2003). In addition, the potential interaction of perennial plants with associated native
411 annual communities in influencing relationships with *Bromus* should be examined. For example,

412 the most competitive native vegetation with *Bromus* could be a competitive perennial species
413 paired with native annuals most competitive with *Bromus*, if such native vegetation types exist.

414

415 **Feasibility of Native Species Treatments.** Since early work in the 1940s and 1950s (Went
416 1942; Muller 1953; Muller and Muller 1956) and with few exceptions such as Brooks (2009),
417 little research has been directed at among-species differences in associations with annual species
418 in southwestern deserts. This early work also concentrated on associations with native annuals,
419 with less emphasis on exotics like *Bromus*. Much contemporary work has further focused on the
420 effects of exotic annuals on native species, rather than which native species most strongly
421 influence exotics. As a result, little research is available to help identify which native species
422 could be amenable to control strategies for *Bromus* and whether competition from native species
423 is a feasible treatment for reducing exotic plants.

424 For competition by native plant species to be an effective treatment for controlling
425 exotics, the native species must: (1) compete with and reduce the exotic, (2) become reliably
426 established via seeding or planting on sites where the exotic grows, and (3) be amenable to
427 propagation and plant material development so that sufficient quantities of the species is
428 available for treatments at operational scales. Native species treatments have shown mixed
429 success at reducing exotic plants, either in the context of oversowing the native into an existing
430 stand of the exotic, or following treatment of the exotic by herbicide or other methods. For
431 example, Bakker and Wilson (2004) found that seeding a mixture of native species reduced the
432 number of plots invaded by the exotic perennial grass crested wheatgrass (*Agropyron cristatum*
433 (L.) Gaertn.) by 33% in Saskatchewan grasslands. Similarly, oversowing the native annual forb
434 firewheel (*Gaillardia pulchella* Foug.) into existing stands of the exotic annual forb

435 bastardcabbage (*Rapistrum rugosum* (L.) All.) reduced productivity of the exotic by up to 72%
436 along Texas roadsides (Simmons 2005). However, neither Thacker et al. (2009) nor Stevens and
437 Fehmi (2009) found that seeding native perennial grasses reduced exotic grasses (including
438 *Bromus tectorum* and buffelgrass [*Pennisetum ciliare* (L.) Link]) of southwestern arid lands. In
439 contrast, Seabloom et al. (2003) found that seeding native perennial grasses into California
440 exotic annual grasslands (including species of *Bromus*) reduced seed production and biomass of
441 the exotics.

442 Further work for evaluating native species treatments is required in deserts where
443 facilitation of exotics by native species appears prevalent (Rodríguez-Buriticá and Miriti 2009).
444 Indeed, we found that *Bromus* had greater growth when grown with natives (compared to when
445 *Bromus* was grown alone) for more than half of the native species in our greenhouse competition
446 experiment, and none of the 16 native perennial species in the field study harbored less cover of
447 *Bromus* than interspaces. These findings suggest that ascertaining native species that at the very
448 least facilitate *Bromus* the slightest may be a reasonable goal for identifying native vegetation
449 types less susceptible to *Bromus* invasion.

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Sources of Materials

452 ¹Las Vegas Springs Preserve, 333 S. Valley View Blvd., Las Vegas, NV 89107.

453 ²National Park Service, Lake Mead National Recreation Area, 601 Nevada Way, Boulder
454 City, NV 89005.

455 ³Kellogg Potting Soil, Kellogg Garden Products Corporate Office, 350 W. Sepulveda
456 Blvd., Carson, CA 90745.

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616 Table 1. Median percent cover of *Bromus rubens* in interspaces and below native perennial
 617 plants (arranged in order of supporting the least to the most *Bromus*) in the eastern Mojave
 618 Desert, southwestern USA. Medians without shared letters differ at $P < 0.05$ (Tukey's test).

Microsite	Median	95% CI ^a	n ^b
Interspace	1 a	1-2	56
<i>Thamnosma montana</i>	2 ab	2-5	22
<i>Bebbia juncea</i>	2 abc	0-9	7
<i>Encelia virginensis</i>	2 abc	1-19	7
<i>Salazaria mexicana</i>	2 abc	2-9	9
<i>Encelia farinosa</i>	3 bc	2-5	30
<i>Coleogyne ramosissima</i>	5 abc	2-5	40
<i>Pleuraphis rigida</i>	5 abcd	2-9	6
<i>Menodora spinescens</i>	5 bc	5-5	37
<i>Psorothamnus fremontii</i>	5 bc	2-9	29
<i>Ambrosia dumosa</i>	5 c	5-9	22
<i>Eriogonum fasciculatum</i>	5 bc	2-38	11
<i>Gutierrezia sarothrae</i>	5 bc	2-9	11
<i>Hymenoclea salsola</i>	9 bcd	2-38	10
<i>Larrea tridentata</i>	9 bcd	2-38	13
<i>Ephedra torreyana</i>	9 c	5-19	28
<i>Krameria erecta</i>	19 d	19-19	37

619 ^aLower and upper bounds of confidence intervals for medians.

620 ^bTotal number measured at seven sites.

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628 Table 2. Summary of the performance of native perennial species with *Bromus rubens* in this
 629 study compared with Brooks (2009), Mojave Desert, southwestern USA. Species were ranked as
 630 poor, medium, or good in their ability for reducing *Bromus* biomass (greenhouse competition
 631 experiment) and for supporting low amounts of *Bromus* in the field (i.e. a ‘good’ rating indicates
 632 that *Bromus* is sparse below a species’ canopy).

Species	This study: competition	This study: field	Brooks (2009): field
<i>Ambrosia dumosa</i>	Medium	Medium	Good
<i>Bebbia juncea</i>	Good	Good	Good
<i>Coleogyne ramosissima</i>	—	Medium	Medium
<i>Eriogonum fasciculatum</i>	Good	Medium	Medium
<i>Hymenoclea salsola</i>	Medium	Poor	Medium
<i>Krameria erecta</i>	—	Poor	Medium
<i>Larrea tridentata</i>	Medium	Poor	Medium
<i>Psoralea fremontii</i>	—	Medium	Poor
<i>Salazaria mexicana</i>	Good	Good	Poor
<i>Thamnosma montana</i>	—	Good	Poor

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644 Fig. 1. Competitive hierarchy of native Mojave Desert species with *Bromus rubens*. Competitive
645 performance is calculated as the percentage of the biomass *Bromus* attained with native species
646 compared to when *Bromus* was grown alone. Low values indicate strong competitive effects of
647 native species on *Bromus*. Error bars are 1 SE of the mean.

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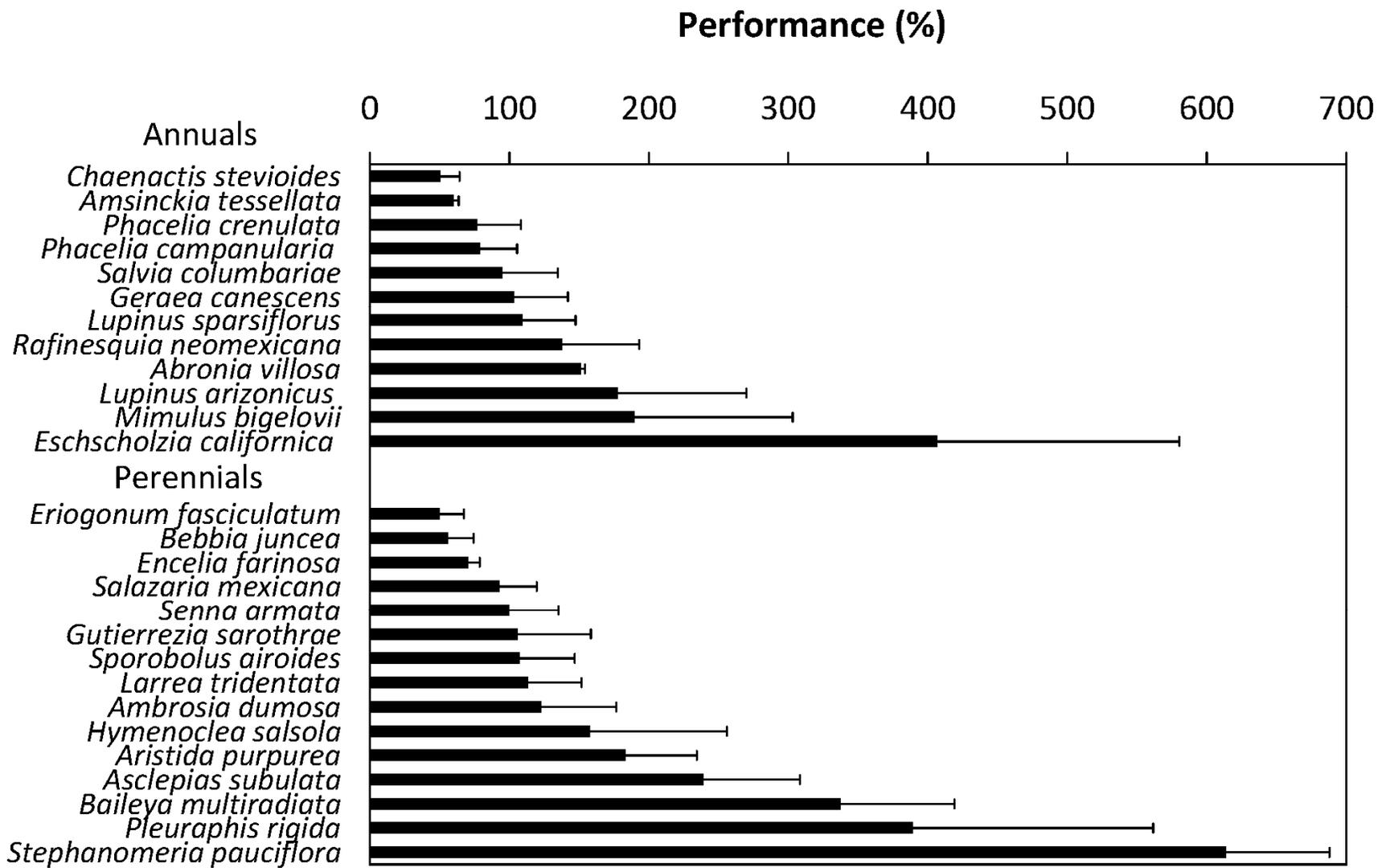
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Interpretive Summary

Identifying native species, for use in revegetation and plant community augmentation projects, capable of reducing the fitness of invasive species is a major thrust in invasive species science and management. We sought to identify native species competitive with *Bromus rubens*, an exotic annual grass increasing fuel loads and facilitating fires devastating to resources in southwestern USA arid lands. In a greenhouse experiment screening the competitive abilities of 27 native species, we found that the native annuals *Chaenactis stevioides* and *Amsinckia tessellata* and the perennials *Eriogonum fasciculatum*, *Bebbia juncea*, and *Encelia farinosa* mostly strongly competed with *Bromus*. In a companion field assessment of *Bromus* below different native perennial plants in the eastern Mojave Desert, *Bromus* cover varied more than 9-fold among 16 species. Species such as *Thamnosma montana*, *B. juncea*, *E. farinosa*, and *E. fasciculatum* showed promise in their ability to support low levels of *Bromus* cover, whereas *Krameria erecta* seems to facilitate *Bromus*. There is potential to develop lists of species for practitioners to recommend or avoid using in desert revegetation projects based on species' interactions with *Bromus*.